



Long-term volatility measurements of submicron atmospheric aerosol in Hyytiälä, Finland

S. A. K. Häkkinen¹, M. Äijälä¹, K. Lehtipalo¹, H. Junninen¹, J. Backman¹, A. Virkkula^{1,2}, T. Nieminen¹, M. Vestenius², H. Hakola², M. Ehn³, D. R. Worsnop^{1,4}, M. Kulmala¹, T. Petäjä¹, and I. Riipinen^{1,5}

¹Department of Physics, University of Helsinki, P.O. Box 64, 00014 Helsinki, Finland

²Finnish Meteorological Institute, Erik Palménin aukio 1, 00560 Helsinki, Finland

³Institute for Energy and Climate Research (IEK-8), Forschungszentrum Jülich, 52425 Jülich, Germany

⁴Aerodyne Research, Inc., Billerica, Massachusetts, USA

⁵Department of Applied Environmental Science and Bert Bolin Centre for Climate Research, Stockholm University, 10691 Stockholm, Sweden

Correspondence to: S. A. K. Häkkinen (silja.hakkinen@helsinki.fi)

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Abstract. The volatility of submicron atmospheric aerosol particles was investigated at a boreal forest site in Hyytiälä, Finland from January 2008 to May 2010. These long-term observations allowed for studying the seasonal behavior of aerosol evaporation with a special focus on compounds that remained in the aerosol phase at 280 °C. The temperature-response of evaporation was also studied by heating the aerosol sample step-wise to six temperatures ranging from 80 °C to 280 °C. The mass fraction remaining after heating (MFR) was determined from the measured particle number size distributions before and after heating assuming a constant particle density (1.6 g cm⁻³). On average 19 % of the total aerosol mass remained in the particulate phase at 280 °C. The particles evaporated less at low ambient temperatures during winter as compared with the warmer months. Black carbon (BC) fraction of aerosol mass correlated positively with the MFR at 280 °C, but could not explain it completely: most of the time a notable fraction of this non-volatile residual was something other than BC. Using additional information on ambient meteorological conditions and results from an Aerodyne aerosol mass spectrometer (AMS), the chemical composition of MFR at 280 °C and its seasonal behavior was further examined. Correlation analysis with ambient temperature and mass fractions of polycyclic aromatic hydrocarbons (PAHs) indicated that MFR at 280 °C is probably affected by anthropogenic emissions. On the other hand, results from the AMS analysis suggested that there

may be very low-volatile organics, possibly organonitrates, in the non-volatile (at 280 °C) fraction of aerosol mass.

1 Introduction

Aerosol particles, i.e. liquid or solid airborne particles, are ubiquitous in the earth's atmosphere. These particles affect our lives by reducing visibility (Cabada et al., 2004), causing adverse health effects (Davidson et al., 2005; Nel, 2005) and affecting the climate (Lohmann and Feichter, 2005; IPCC, 2007). Aerosol particles reflect and absorb incoming solar radiation (Bohren and Huffman, 1983) and affect water circulation, precipitation and the earth's radiative balance via cloud formation (Ramanathan et al., 2001; Rosenfeld et al., 2008; Spracklen et al., 2008). Aerosol particles are either directly emitted into the atmosphere in the condensed or solid phase, for example pollen from vegetation or soot from combustion, or formed in the atmosphere from ambient vapors (Kulmala et al., 2004).

Volatility, i.e. the tendency of a molecule to evaporate from the aerosol phase, as compared with its flux to the particles, is among the most important factors determining if the molecule partitions into the aerosol phase. Atmospheric aerosols are typically complex mixtures of inorganic and organic compounds. The identities and thermodynamic properties of inorganic aerosols are relatively well-known – at least

compared with the organic fraction of atmospheric aerosol. Ambient air contains thousands of organic compounds that have complex structures and formation mechanisms (Goldstein and Galbally, 2007). A large number of these compounds still remain unidentified. Depending on the location and season, 20–90 % of the aerosol mass consists of organics (Kanakidou et al., 2005; Zhang et al., 2007; Jimenez et al., 2009). A large fraction of the organics found in aerosol particles are semi-volatile under atmospheric conditions (Robinson et al., 2007), meaning that they can be present in both gaseous and particulate phases. In order to yield reliable estimates for the total aerosol budget and to understand how the semi-volatile compounds affect the formation and fate of atmospheric aerosols, the volatility of the aerosol constituents needs to be known. Information on the aerosol volatility also gives indirect indications of the aerosol composition.

The volatility of atmospherically relevant aerosol particles has recently been investigated extensively both in the laboratory (An et al., 2007; Faulhaber et al., 2009; Cappa and Wilson, 2011; Saleh et al., 2011) and in the field. Field measurements have been performed in various environments: in rural and urban background areas (O'Dowd et al., 2000; Wehner et al., 2005; Ehn et al., 2007; Backman et al., 2010; Lee et al., 2010) as well as in polluted cities and roadsides (Wehner et al., 2004; Huffman et al., 2009; Birmili et al., 2010; Tiitta et al., 2010). In these studies, the general principle for measuring aerosol volatility has been similar: aerosol samples have been heated in a thermodenuder and the size (diameter, volume or mass) of aerosol particles determined before and after heating. Thermodenuders (TD) come in different designs, and the detected volatility depends on the temperature, residence time and total aerosol loadings in the TD (Cappa et al., 2010; Riipinen et al., 2010; Saleh et al., 2011; Fuentes et al., 2012) – making direct quantitative comparison of different volatility measurements challenging. One of the challenging aspects of the interpretation of TD data is the fact that time that it takes for the aerosol to reach equilibrium in a TD depends on the aerosol properties (Cappa et al., 2010; Riipinen et al., 2010; Saleh et al., 2011; Fuentes et al., 2012). This means that the ambient/laboratory generated aerosol does not necessarily achieve equilibrium in the currently available TDs – depending on the aerosol residence time in the TD (An et al., 2007; Riipinen et al., 2010; Saleh et al., 2011). Commercially available thermodenuders (TOPAS GmbH by TSI Inc, MN; Dekati TD by Dekati Ltd, Tampere, Finland) and the ones developed by Bartscher et al. (2001) and Wehner et al. (2002) have residence times of a few seconds. For measurements of semi-volatile organics TDs with longer residence times have been developed recently (An et al., 2007; Huffman et al., 2008). The TD used in this study is the same as described in Ehn et al. (2007) with a maximum heating temperature of 280 °C and a residence time around a second.

Atmospheric aerosols are typically a mixture of compounds with different volatilities. For individual compounds

at a fixed residence time, the evaporation inside a thermodenuder typically happens around a specific temperature, while for complex atmospheric mixtures the temperature-dependence of the aerosol remaining after heating is smoother (e.g. Riipinen et al., 2010). A large fraction of atmospheric organics and many inorganic aerosol compounds evaporate at temperatures below 300 °C (Turpin et al., 2000; Raatikainen et al., 2010), for example ammonium sulfate at around 200 °C and ammonium nitrate at around 60–75 °C even with residence times less than a second (Johnson et al., 2004; Philippin et al., 2004; Villani et al., 2007). However, some compounds, such as black carbon (BC), sea salt and crustal material, are not volatilized even at 300 °C.

Almost all the volatility observations of submicron aerosol particles, even of nucleated particles around 3–60 nm in diameter, to date have found a non-volatile residual in the particles that does not evaporate at temperatures around 280–300 °C in similar setups as the one used in this study (Wehner et al., 2005; Ehn et al., 2007; Birmili et al., 2010). In this study the term “non-volatile” refers to material not evaporated at 280 °C. In polluted areas, the observed non-volatile aerosol material has been found to have a clear correlation with light-absorbing aerosol (Birmili et al., 2010). Kalberer et al. (2004) show that given enough time atmospheric organic compounds may form low-volatile polymers. Recently Smith et al. (2010) reported that aminium salt formation takes place in atmospheric nanoparticles and that these low-volatile salts contribute significantly to the growth of nanoparticles in various environments, also in boreal forests. Barsanti et al. (2009) suggested that amines may have a noticeable role in the formation of organic salts. All of the aforementioned processes may be linked to the observations of low-volatility material in the small atmospheric particles.

The evaporation properties of atmospheric submicron aerosol particles were investigated in this study by examining the mass reduction of aerosols due to evaporation at six temperatures ranging from 80 °C to 280 °C. Measurements were performed continuously during a period of two and a half years (January 2008 to May 2010) at the boreal forest site of Hyytiälä, Finland. Compared to previous studies performed in Hyytiälä (e.g. Ehn et al., 2007; Raatikainen et al., 2010), this study is the first one focusing on long-term variation of aerosol evaporation. The volatility data was compared to BC measurements, meteorological data and trajectory analysis as well as to aerosol composition data from filter analysis and aerosol mass spectrometer (AMS). Our main aims were to (1) study the temperature dependence of aerosol volatility in Hyytiälä, (2) investigate the seasonality of aerosol volatility and its connection to other atmospheric variables, and (3) examine the chemical composition of the non-volatile fraction (not evaporating at 280 °C) of atmospheric sub-micron particles, particularly to answer the question of how large fraction of this very low-volatility material was BC vs. other compounds.

2 Materials and methods

2.1 Measurement site

Measurements for this study were performed at a rural background measurement station SMEAR II (Station for Measuring Ecosystem-Atmosphere Relations II) located in Hyytiälä (61°51' N, 24°17' E, 181 m a.s.l.), in Southern Finland (Vesala et al., 1998; Kulmala et al., 2001; Hari and Kulmala, 2005). The SMEAR II station is surrounded by 48 year old boreal conifer forest that covers an area spanning tens of kilometers to North and North-East. The nearest large cities are Tampere (around 60 km to South-West, 213 000 inhabitants) and Jyväskylä (around 100 km to North-East, 131 000 inhabitants). Ambient meteorological conditions such as temperature, relative humidity, solar radiation, wind speed and direction as well as concentrations of several trace gases, such as carbon dioxide (CO₂), carbon monoxide (CO), sulfur dioxide (SO₂) and nitrogen oxides (NO_x), are measured continuously at the station. Also the fluxes of compounds such as water, ozone and several volatile organics are measured. In addition, measurements of aerosol properties such as aerosol particle number size distributions have been ongoing at the station continuously since 1996 (Mäkelä et al., 1997).

2.2 Instrumentation

2.2.1 DMPS

The ambient aerosol number size distribution was measured from 1 January 2008 to 24 May 2010 (see Fig. 1) with a twin-DMPS (Differential Mobility Particle Sizer, Hoppel, 1978; Aalto et al., 2001) consisting of two separate DMPSs operated in parallel. The first DMPS determines the size distribution of small particles, 3–50 nm in diameter, and the other measures larger particles, 15 nm–1 µm in size. Together they cover the size range of 3–1000 nm. Both DMPSs sheath flows are maintained using closed loop circulation (Jokinen and Mäkelä, 1997) and at a relative humidity below 20 %. Condensation particle counters (CPC, Stolzenburg and McMurry, 1991) used for particle counting are TSI-3025 and TSI-3010 (TSI Inc.), respectively. The time resolution of the combined system is 10 min.

2.2.2 VDMPS

The evaporation of submicron aerosol particles was investigated using a volatility differential mobility particle sizer (VDMPS), which consists of two separate parts, a thermodenuder (TD, Ehn et al., 2007) and a DMPS. The thermodenuder heats up the sample air, after which the remaining aerosol material is directed to the DMPS system. The heating temperature can be either held constant or scanned over several temperatures. The VDMPS used in this study measured the number size distribution of heated particles of 20–

1000 nm in diameter. A TSI model 3010 CPC was used for particle counting. The thermodenuder was a 50 cm long tube fabricated from stainless steel with outer radius of 10 mm. The average residence time, i.e. the time the aerosol sample spends in the oven, was 1.2 s leading into a minimum residence time of 0.6 s at the centerline. The time resolution of the VDMPS was 10 min. No absorptive material was used in the thermodenuder.

The VDMPS and DMPS systems were located in the same measurement cottage and they used the same total aerosol inlet. No PM_{1.0} cyclone was used in front of the VDMPS, so the VDMPS was occasionally affected by evaporated coarse mode particles (> 1 µm) not detected by the DMPS system. These particles were detected in VDMPS after heating to high temperatures and caused noise in the VDMPS data at the larger end of particle sizes. In order to remove this noise, the size range of 20–500 nm was selected for further analysis.

The VDMPS was temperature-scanning from 1 January 2008 to 27 April 2009. The TD heated the aerosol sample to six temperatures, 80 °C, 120 °C, 160 °C, 200 °C, 240 °C and 280 °C with a time resolution of about one hour. From July 2009 onwards the aerosol sample was heated to a constant temperature, 280 °C (Fig. 1) to obtain a complete time series of evaporation at a single temperature. Total data coverage for the measurement period (1 January 2008 to 24 May 2010) was about 75 %.

The major loss processes in the heating tube are caused by thermophoresis and Brownian diffusion of small particles (Wehner et al., 2002). Due to these losses, VDMPS raw data underestimates the number concentration of submicron particles. Errors in particle number concentrations were corrected using experimentally derived particle losses presented by Ehn et al. (2007). The losses for aerosol particles above 15 nm in diameter were observed to be around 20 % of the total particle number concentration when heated to 280 °C. The same correction fit for particle penetration through the oven was used for all the six heating temperatures. The error resulting from this has been estimated to be only a few percent (Wehner et al., 2002).

2.2.3 Aethalometer

An aethalometer (Hansen et al., 1984) was used to measure the mass concentration of BC after a PM_{2.5} inlet. In the data inversion it is assumed that the BC mass concentration and light attenuation are linearly related, and BC mass is calculated from the rate of change of light transmission through a filter at seven wavelengths. However, the presence of other light-scattering and absorbing particles, such as sulfate or brown carbon, on the filter can cause non-linearities in the relationship between the light transmission and the BC concentration (Reid et al., 1998; Bond et al., 1999). The aethalometer signal needs to be corrected for these effects (e.g. Weingartner et al., 2003; Arnott et al., 2005; Andreae and Gelencsér, 2006; Virkkula et al., 2007). The errors caused by

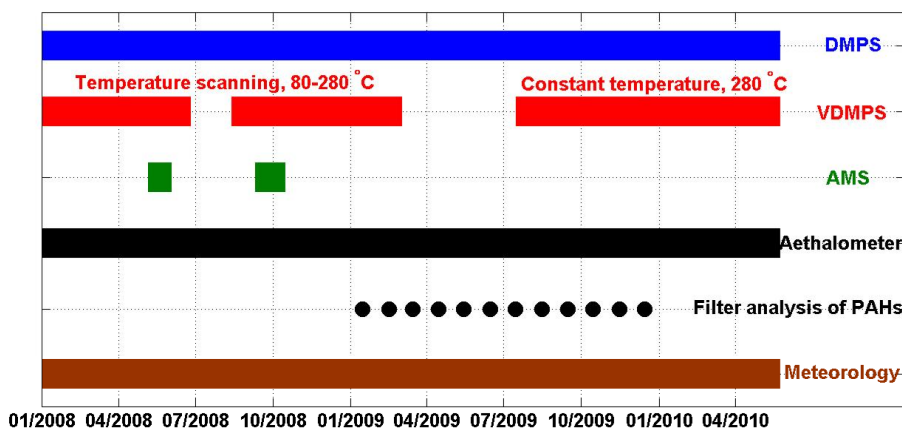


Fig. 1. Instruments used in this study and their functioning times from January 2008 to May 2010.

brown carbon can be reduced by using longer wavelengths, most preferably 880 nm (Weingartner et al., 2003).

Aethalometer data was corrected using the approach introduced by Weingartner et al. (2003), the resulting BC concentrations (M_{BC}) being defined as

$$M_{BC} = \frac{M_{BC,raw} \sigma_{abs,0}(\lambda)}{\sigma_{abs}(\lambda) C(\lambda) R(ATN)}, \quad (1)$$

where $M_{BC,raw}$ is the BC concentration measured with an aethalometer (non-corrected), $\sigma_{abs,0}$ is the wavelength-dependent absorption mass cross-section used by the manufacturer of the aethalometer, σ_{abs} is the absorption mass cross-section calculated using the absorption mass cross-section of MAAP (Multiangle Absorption Photometry, Petzold and Schönlinner, 2004) as a reference point ($6.6 \text{ m}^2 \text{ g}^{-1}$ at 637 nm), and C and $R(ATN)$ are the calibration parameters. The calibration factor $R(ATN)$ can be written as

$$R(ATN) = \left(\frac{1}{a(1 - \omega_0) + 1} - 1 \right) \frac{\ln(ATN) - \ln(10\%)}{\ln(50\%) - \ln(10\%)} + 1 \quad (2)$$

where a is 0.85 at 660 nm (Weingartner et al., 2003), ATN is light attenuation through the filter and ω_0 is a single-scattering albedo which can be derived from particle absorption and scattering measurements. Aerosol light scattering measurement data from a nephelometer (Anderson et al., 1996) was available for the period of interest. Values for absorption mass cross-sections and C factors are presented in Table A1 in Appendix A. The C factors were calculated by using the average C value determined from Hohenpeisenberg MAAP-data by Collaud-Coen et al. (2010). Wavelength dependence of the C factor was obtained from Arnott et al. (2005). Corrected BC data at 880 nm wavelength with a time resolution of five minutes was used in the analysis.

2.2.4 AMS

The Aerosol Mass Spectrometer (AMS) used in this study was an Aerodyne Compact Time-of-Flight Aerosol Mass

Spectrometer (C-ToF AMS, Aerodyne Research Inc., Billerica, US), which measures the chemical composition of submicron particles ($< 600 \text{ nm}$) and has a 50 % detection efficiency cutoff at 600 nm (Jayne et al., 2000; Drewnick et al., 2005; Liu et al., 2007).

The C-ToF AMS features an aerodynamic lens for concentrating the particles into a narrow beam, a particle time-of-flight (PTof) chamber for particle size distribution measurement, thermal vaporization of the sample particles in a 600°C tungsten oven, 70 eV electron impact ionization (EI) of the produced vapor, and a compact ion-time-of-flight mass spectrometer (C-ToF MS) to obtain a mass spectrum of the ions. The “hard” EI ionization method provides quantitative, yet highly fragmented ion spectra from the parent molecules.

The mass spectra of ions is interpreted, using knowledge of EI molecular fragmentation patterns, to assign the aerosol total mass loading into subgroups of different chemical composition: sulfates, organics, nitrates, ammonium and chlorides. The individual mass-per-charge signals and their relative ratios can be used to further characterize the composition of the sample aerosol. The final data is scaled with a collection efficiency factor (CE) to compensate for losses inside the AMS.

Because of the thermal vaporization method, the AMS does not quantitatively see “refractory compounds”, i.e. those that are not vaporized below 600°C , such as sea salt and BC. Particulate water is also not quantitatively measured and is not included in the AMS-measured total aerosol mass.

Data used in this study was collected during two month-long AMS-campaigns: one in late spring 2008 and one in early fall 2008 (Fig. 1).

2.2.5 Meteorological parameters and PAH-data

Meteorological parameters as well as mass concentrations of polycyclic aromatic hydrocarbons (PAH) were also measured at SMEAR II. Ambient temperature and solar radiation were continuously recorded at a 72-m-tall mast at different

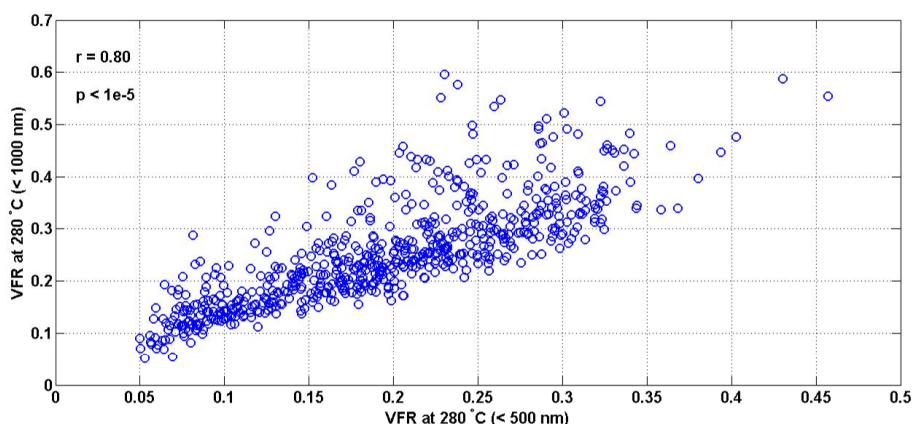


Fig. 2. Correlation between the VFR (DMPS and VDMPS < 1000 nm) at 280 °C and the VFR (DMPS and VDMPS < 500 nm) at 280 °C. Correlation coefficient (r) and p -value describing the significance of the correlation are provided in the figure.

heights. In this study the radiation data was acquired from the height of 18 m whereas the ambient temperatures were measured at a lower measurement stage (4.2 m). The meteorological data was averaged over 30 min. Mass concentrations of PAHs were analyzed from filter samples (PM₁₀) in Hyytiälä during 2009 (see Vestenius et al., 2011) (Fig. 1).

3 Data analysis

3.1 Volume fraction remaining

The evaporation of submicron particles was studied by comparing the ambient and heated particle number and volume size distributions measured by the DMPS and VDMPS. The particle volume size distribution was calculated from the number size distribution data assuming spherical particles. The volume fraction of aerosol particles remaining after heating was then defined by comparing the total volume of heated aerosol ($V_{\text{tot,VDMPS}}$) to the total volume of ambient aerosol ($V_{\text{tot,DMPS}}$). Volume fraction remaining (VFR) can be written as

$$\text{VFR} = \frac{V_{\text{tot,VDMPS}}}{V_{\text{tot,DMPS}}}. \quad (3)$$

As mentioned above (see Sect. 2.2.2), the size range of 20–500 nm was used for both instruments due to occasional noise in the VDMPS data at larger particle sizes. Figure 2 illustrates the correlation between VFR (DMPS and VDMPS < 1000 nm) at 280 °C and VFR (DMPS and VDMPS < 500 nm) at 280 °C. The correlation is clear ($r = 0.80$) and therefore confirms that the size range of 20–500 nm is representative of the total aerosol evaporation behavior. Absolute minimum estimates for the VFRs in different temperatures can be calculated selecting different size ranges for the calculations (VDMPS < 500 nm and DMPS < 1000 nm), resulting in VFR values around 20 % smaller than presented here.

We assumed a constant particle density of 1.6 g cm^{-3} throughout the aerosol size distribution based on the work of Kannosto et al. (2008), who report a range of $1.1\text{--}2 \text{ g cm}^{-3}$ for accumulation mode particle densities in Hyytiälä. The mass fraction remaining after heating (MFR) was thus assumed to be equivalent to VFR in terms of density. The assumption of constant particle density naturally leads to uncertainties in the MFR if the particle density is in fact changing with size. This uncertainty is likely higher at the higher heating temperatures, because of the bigger change in the particle chemical composition due to evaporation. If the particles are denser after the heating than before, the real MFR values are higher values than presented in this study and vice versa. If the non-volatile residual at 280 °C would consist of BC only (density of 1.8 g cm^{-3} , McMeeking et al., 2010), the MFR at 280 °C would be biased high, with around 13 % as compared with the reported values. Similarly, if the particle density after the heating would be smaller than assumed, e.g. 1 g cm^{-3} , there would be a 38 % low bias in the MFR at 280 °C.

3.2 Black carbon fraction

The maximum fraction of BC in the aerosol mass (BCF) was defined as

$$\text{BCF} = \frac{M_{\text{BC}}}{M_{\text{tot,DMPS}}}, \quad (4)$$

where M_{BC} is the mass concentration of BC from the aethalometer data (see Eq. 1) and $M_{\text{tot,DMPS}}$ is the total mass concentration of ambient 20–500 nm aerosol as determined from the DMPS data. To investigate the maximum contribution of BC to the MFR at 280 °C, all BC mass was removed from the observed MFR. Mass fraction remaining at 280 °C without the effect of BC is then defined as follows

$$\text{MFR}_{\text{non-BC}} = \text{MFR} - \text{BCF} = \frac{M_{\text{tot,VDMPS}} - M_{\text{BC}}}{M_{\text{tot,DMPS}}}. \quad (5)$$

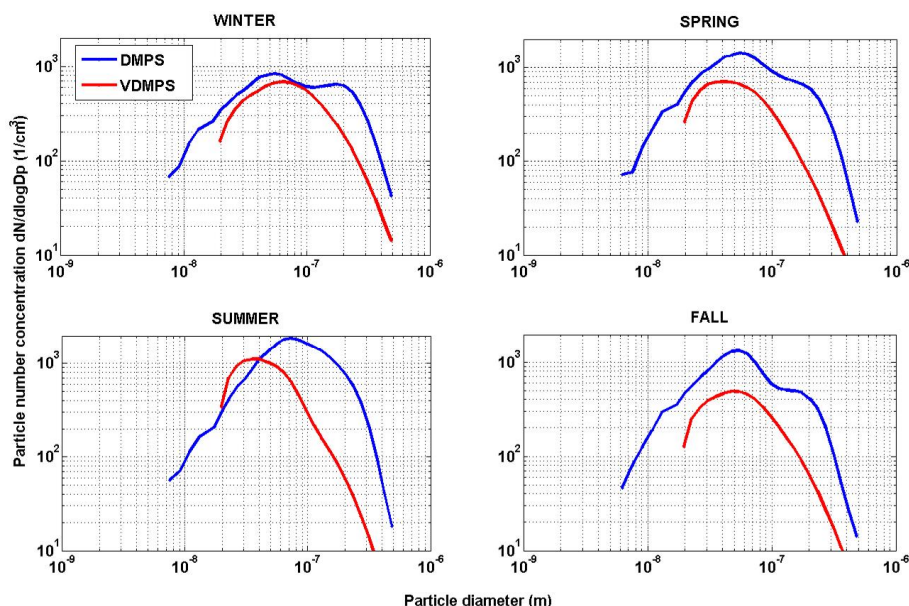


Fig. 3. Median particle number size distributions from DMPS and VDMPS at 280 °C in different seasons.

It is now important to note that the aethalometer determined the BC mass concentration of PM_{2.5} particles. The BCF, however, was calculated assuming that all the BC mass is distributed to particles smaller than 500 nm. Hitzengerger and Tohno (2001) estimate that close to urban areas approximately 15–20 % of the BC mass is in particles larger than 500 nm. If the same is true also in Hyytiälä the error introduced to the MFR_{non-BC} at 280 °C would increase the value of MFR_{non-BC} with a factor of 1.3 at maximum.

3.3 Air mass trajectories

To investigate the relationship between atmospheric aerosol volatility and air mass properties, the air mass origin was determined with trajectory analysis. Back trajectories of air masses arriving at Hyytiälä were modeled with HYSPLIT (HYbrid Single-Particle Lagrangian Integrated Trajectory model, Draxler and Hess, 1997), which runs via internet (Draxler and Rolph, 2011). We used hourly air mass trajectories at a 100 m arrival height above ground level. The trajectories were calculated 96 h backwards in time. In addition to the trajectories, HYSPLIT estimates the height of the mixing (boundary) layer. The average direction of the arriving air mass was determined using the trajectories that were below twofold the mixing layer height.

4 Results and discussion

4.1 Ambient and non-volatile number and volume size distributions

Figure 3 shows the median ambient and heated (280 °C) number size distributions for winter (December–February), spring (March–May), summer (June–August) and fall (September–November). The median ambient number size distributions were bi- or trimodal with Aitken and accumulation modes. The highest number concentrations were observed in the Aitken mode, and during summer and spring months the total particle number concentration was the highest – in agreement with previous observations (Dal Maso et al., 2005). Also nucleation mode (3–25 nm) particles are frequently observed in Hyytiälä. This mode cannot be clearly observed in Fig. 3 because of the averaging, as it has a distinct diurnal pattern. However, when examining aerosol number size distributions for days and nights separately, the nucleation mode was visible in day time number size distributions, especially during spring and fall months.

The median heated (280 °C) number size distributions were monomodal (Fig. 3). In general, the particle modes moved towards smaller sizes when heated, and the bimodal structure disappeared. Aerosol total number concentration after heating was noticeably lower than in the ambient air. There are at least two possible reasons for this, after the temperature-dependent losses are accounted for. First, a large fraction of the smallest aerosols can evaporate below the detection limit of the VDMPS system (20 nm). Second, if the aerosol is an external mixture of more and less volatile components, some particles can evaporate completely while

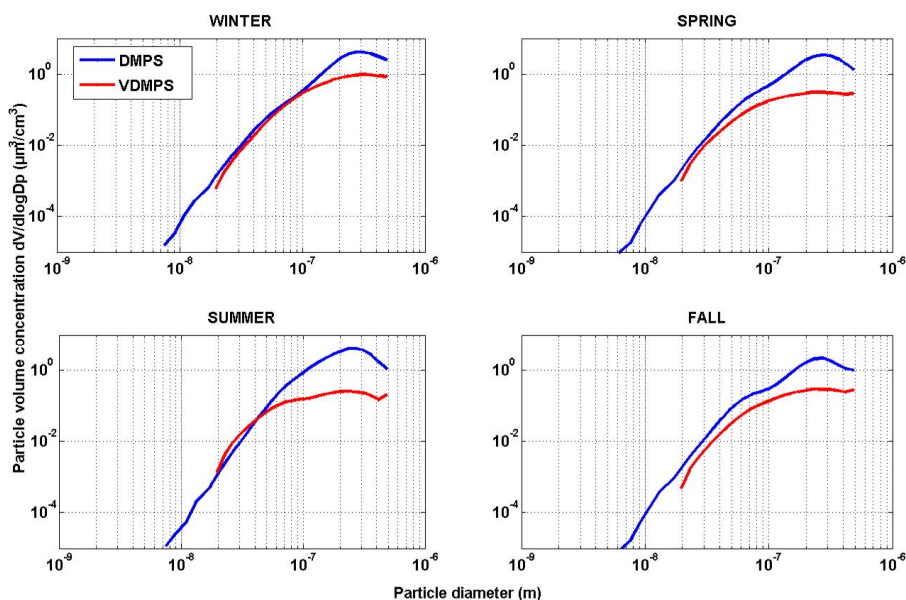


Fig. 4. Seasonal median volume size distributions from DMPS and VDMPS at 280 °C.

others remain close to their original size. We expect the former effect to be more important, in accordance with earlier findings where all nucleation mode particles were found to contain material that did not evaporate at temperatures of 280 °C (Ehn et al., 2007).

Figure 4 illustrates the median volume size distributions of particles in ambient conditions and after heating to 280 °C during different seasons. While the ambient and heated number and volume size distributions for the winter, spring and fall look relatively similar, the corresponding distributions for summer differ from the other seasons. During the summer the distinction between the ambient Aitken and accumulation modes is less clear and the average diameter of the heated aerosol smaller than during the other seasons.

4.2 Temperature-dependence of MFR

Figure 5 shows the time series of the MFR at different TD temperatures. The aerosol particles were more volatile during summer, whereas the highest values of the MFR occurred during winter months (Table 1). On average 19 % of the aerosol mass was left after heating the aerosol to 280 °C. In general the temporal variation of the MFR values was relatively similar, independent on the TD temperature (see Fig. 5). During the spring (especially in April/May 2008 and in March 2009), however, the MFRs below 200 °C seemed to show an increasing trend while the MFRs at higher temperatures were decreasing. This indicates an increase in the relative fraction of aerosol components that volatilize around 200 °C during the spring, as compared with the other months. These observations are in line with the results reported by Raatikainen et al. (2010), who studied the connection between aerosol composition and volatility during spring 2005

Table 1. Median total particle volume concentrations measured by the DMPS and VDMPS in different seasons as well as the VFR (MFR) of 20–500 nm particles at 280 °C.

Season	Instrument	Total particle volume concentration ($\mu\text{m}^3 \text{cm}^{-3}$)	VFR
Winter	DMPS	1.72	0.25
	VDMPS	0.38	
Spring	DMPS	1.75	0.17
	VDMPS	0.25	
Summer	DMPS	1.99	0.09
	VDMPS	0.18	
Fall	DMPS	1.00	0.20
	VDMPS	0.18	

in Hyttälä. They noted an anticorrelation between the MFR values between 50–150 °C and 150–280 °C, and concluded that the material volatilizing around 150 °C was mainly organic, while the mass remaining at the higher temperatures consisted of a mixture of organic and inorganic compounds. The MFR at 280 °C was generally slightly higher at night (0.21) as compared to daytime values (0.19).

Figure 6 shows the temperature dependence of the MFR of this study and a selection of other relevant studies. Our results are in a qualitative agreement with the earlier research. However, it should be noted that residence times and aerosol type varied from study to study. These parameters affect strongly the evaporation of aerosol particles. Ehn et al. (2007) and Wehner et al. (2005) investigated the volatility of nucleation mode particles in remote background sites

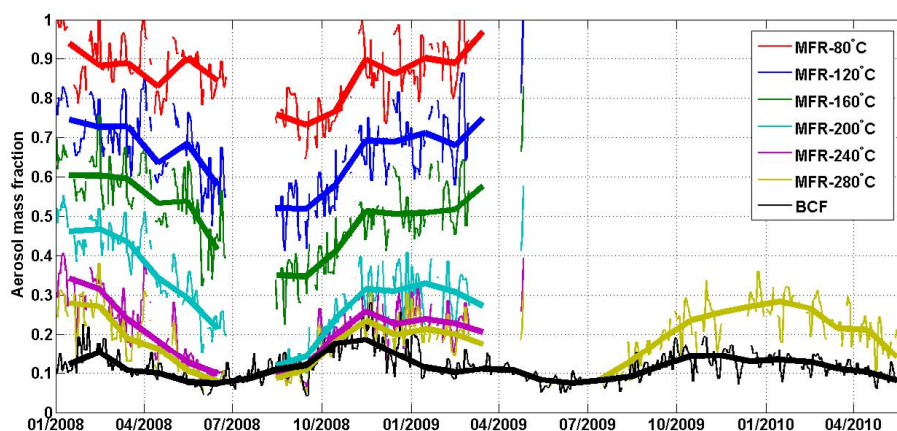


Fig. 5. Aerosol mass fraction remaining (MFR) after heating to six temperatures from 80 °C to 280 °C is presented (colored lines). The black line describes the maximum mass fraction of BC (BCF) in aerosol particles. Thicker lines show the monthly pattern of the parameters presented.

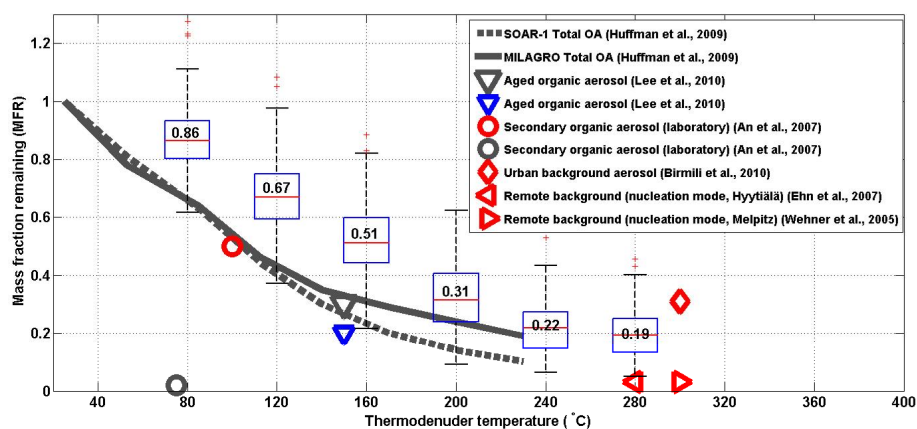


Fig. 6. Comparison of thermograms between this study (the boxplot) and previous volatility studies. Coloring describes the residence times used (red: < 2 s, grey: 10–16 s, blue: > 100 s). SOAR-1 and MILAGRO campaign results are adapted from Huffman et al. (2009). Lee et al. (2010) investigated the volatility of aged organic aerosol during FAME-2008 campaign in Finokalia and volatility measurements of urban background aerosol were performed by Birmili et al. (2010) in Germany. Ehn et al. (2007) and Wehner et al. (2005) measured the volatility of nucleation mode particles in Hyytiälä and in Melpitz, respectively. Measurement results of laboratory generated secondary organic aerosol from An et al. (2007) are also presented.

by examining the change in aerosol mean diameter when heated. Other studies, including our study, determined the volume/mass decrease of particles.

In general the aerosol volatility measured in this study seems to fall close to the lower end of the literature values – although the different residence times and total aerosol loadings make the direct comparison challenging. More aerosol material is expected to evaporate when longer residence times are used. This can be observed when aerosol is heated with varying residence times e.g. in An et al. (2007) and Lee et al. (2010). During the SOAR-1 and MILAGRO campaigns in polluted urban areas in California and Mexico City, volatility of organic aerosol (OA) was investigated at several temperatures up to 230 °C (Huffman et al., 2009). Slopes of these thermograms are very similar to the slope of our volatil-

ity thermogram. The studies by Birmili et al. (2010), Ehn et al. (2007) and Wehner et al. (2005) used similar VDMPS systems with < 2 s residence time as the one used in this study to investigate the volatility of atmospheric aerosol size distributions around 280–300 °C. Our results fall in between these studies, indicating slightly a higher volatility than the urban study by Birmili et al. (2010), but lower than those of Ehn et al. (2007) and Wehner et al. (2007). It should be borne in mind though that Ehn et al. (2007) and Wehner et al. (2007) both concentrated on the volatility of the nucleation mode instead of the total aerosol mass.

During recent years, a number of studies have pointed out challenges in quantitative interpretation of TD or VDMPS data (Cappa et al., 2010; Riipinen et al., 2010; Saleh et al., 2011; Pierce et al., 2011) for compounds – in particular

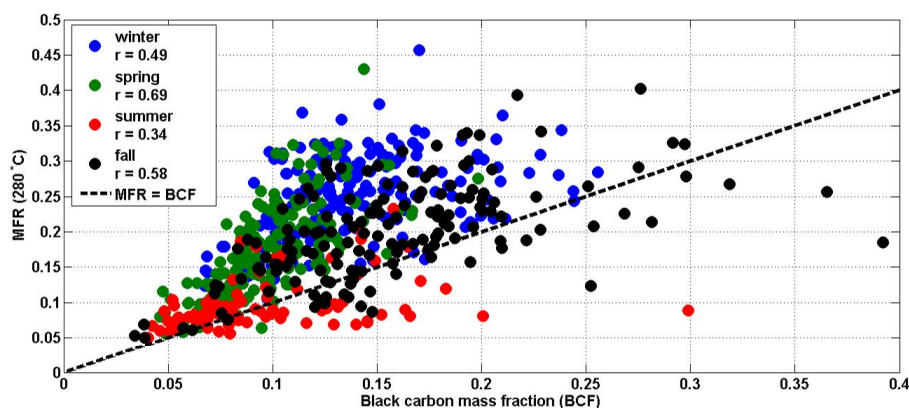


Fig. 7. Mass fraction remaining (MFR) at 280 °C as a function of BC mass fraction during different seasons. A clear correlation is observed during winter, spring and fall months (p -value $< 10^{-5}$). The black line describes the situation where the entire non-volatile aerosol mass (at 280 °C in 1.2 s) may be due to BC. The existence of measurement points below this line most likely arise from the attribution of all $\text{PM}_{2.5}$ BC to sub-500 nm aerosol.

mixtures – whose properties are not well-known. Taking into account these uncertainties in the analysis of the VDMPS data, we do not intend to give a quantitative answer on the variation of the thermodynamic properties of the aerosol in Hyytiälä based on the calculated MFR values. However, even accounting for the uncertainties it can be concluded that the material that does not evaporate at temperatures higher than 280 °C has likely a very low saturation vapor pressure. As an example, using the approach presented by Riipinen et al. (2010), a two-component system consisting of a non-volatile fraction (BC) and a semi-volatile organic fraction (accommodation coefficients of unity, particle diameter 100 nm and mass concentration of $3.0 \mu\text{g m}^{-3}$) would require an organic vaporization enthalpy of 64 kJ mol^{-1} and saturation vapor pressure of 10^{-6} Pa (saturation concentration of $0.15 \mu\text{g m}^{-3}$ or $6 \times 10^8 \text{ cm}^{-3}$ assuming vapor molecule mass 150 amu) or lower to reproduce the temperature-dependence of the VDMPS data (see Fig. 6) and to not fully evaporate at 280 °C over the time of a second (see also Pierce et al., 2011). The obtained vaporization enthalpy value is too low to represent a realistic single-component organic aerosol. However, previous studies have shown that multicomponent organic aerosols (such as probably present in our study) often show a much shallower temperature-dependence than can be represented with a single-component vaporization enthalpy and saturation vapor pressure (e.g. Riipinen et al., 2010). It is also possible that particles react in the heated tube and form compounds with lower volatilities. In any case, the observed non-volatile residual at 280 °C gives indirect information on the chemical composition of the very low-volatile material in aerosol particles.

4.3 Contribution of BC to MFR

Although located in a rural remote area, polluted air masses transport BC to Hyytiälä. Seasonal behavior of the maxi-

mum BC fraction in aerosol particles (see Eq. 4) is presented in Fig. 5. The BCF was the highest during winter and fall, around 13–15 %. In summer and spring months around 8–10 % of the total aerosol mass consisted of BC.

The positive correlation observed between the BCF and MFR at 280 °C (Fig. 7) supports the assumption that BC is connected to the least-volatile fraction of atmospheric aerosol. A similar observation was made by Birmili et al. (2010) in an urban environment. Even though the BCF was the highest during winter and fall (see also Hienola et al., 2012), the best correlation (Pearson correlation) between the BCF and the MFR at 280 °C was observed during spring ($r = 0.69$, p -value $< 10^{-5}$). However, during winter and fall also the MFR at 280 °C was the highest, which may have affected the observed correlations. In general the correlation in summer was the lowest ($r = 0.34$, p -value 10^{-3}), although specifically in summer the BC concentrations were often of similar magnitude than the concentrations of the material that remained non-volatile at 280 °C. This suggests that in the summer a relatively larger fraction of the BC is in particles larger than 500 nm in diameter as compared with the cold months. This is reasonable, since in the winter there are more local sources of BC producing relatively smaller particles than in the summer when most of the BC is likely due to long-range transport. Depending on season BC explained 55–87 % (median values) of the non-volatile residual. The lowest contribution of BC was observed during winter and the highest during summer. In most cases the non-volatile aerosol mass at 280 °C could not be explained by BC only. A similar finding was reported by Backman et al. (2010) who investigated the behavior of particle single-scattering albedo after heating in an urban background area in Helsinki, Finland. They concluded that BC alone could not explain the concentrations of material remaining non-volatile at 280 °C. Neither crustal material nor sea salt is a significant contributor to

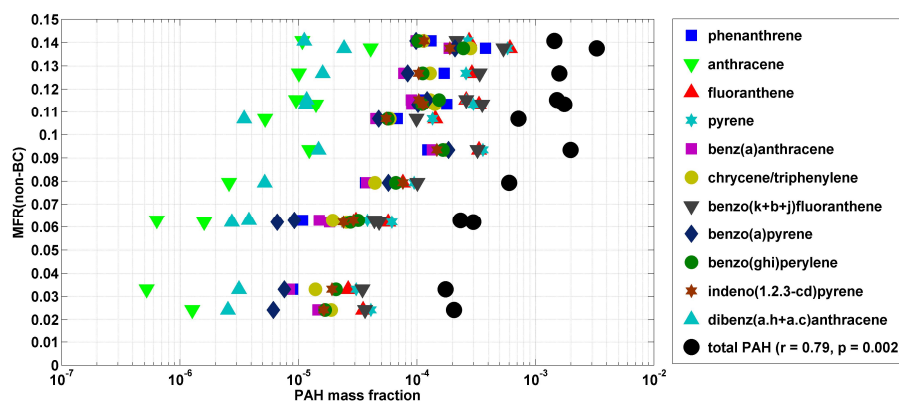


Fig. 8. Aerosol mass fraction remaining ($MFR_{\text{non-BC}}$) at 280 °C as a function maximum mass fractions of PAHs. Monthly means have been used. Note the logarithmic scale.

submicron aerosol composition in Hyytiälä area (Saarikoski et al., 2005). To get additional information on the chemical composition of the least-volatile aerosol fraction, $MFR_{\text{non-BC}}$ at 280 °C was correlated with atmospheric parameters as well as with aerosol chemical composition data obtained from filter analysis and from the AMS.

4.4 Dependence of $MFR_{\text{non-BC}}$ on environmental variables and PAHs

Over the whole measurement period the $MFR_{\text{non-BC}}$ at 280 °C was higher at lower ambient temperatures ($r = -0.46$ with p-value of $< 10^{-5}$). Solar radiation did not affect the $MFR_{\text{non-BC}}$ at 280 °C as strongly as the ambient temperature ($r = -0.27$ with p-value of $< 10^{-5}$). This also indicates that the effect of the height of the boundary layer is not as significant as the ambient temperature. In cold conditions the ambient aerosol mixture is different compared to warm air masses due to different vapor and particle sources. The need for e.g. residential heating during winter is high which indicates that there are more very low-volatile anthropogenic compounds in the atmosphere. The SMEAR II sampling lines are at room temperature and aerosol spends more than 5 s in these lines before arriving at the TD. This should be sufficient for the sample to reach room temperature before being heated to the TD temperature. In addition, the temperature of the aerosol flow inside the TD was measured at the centerline. This suggests that the dependence of the $MFR_{\text{non-BC}}$ at 280 °C on the ambient temperature is not likely to be due to e.g. residence time related artifacts in the aerosol sampling.

The correlation between atmospheric polycyclic aromatic hydrocarbons (PAHs), emitted from combustion processes (Anttila et al., 2005; Rissanen et al., 2006), and the $MFR_{\text{non-BC}}$ at 280 °C was investigated. The monthly median values of $MFR_{\text{non-BC}}$ at 280 °C in sub-500 nm particles as a function of the maximum mass fraction of PAHs are presented in Fig. 8. It can be clearly seen that the particles were relatively more non-volatile with high PAH mass frac-

tions. The correlation coefficient between the $MFR_{\text{non-BC}}$ at 280 °C and the sum of PAH-fractions was as high as 0.79 with a p-value of 0.002 (Fig. 8). The mass concentrations of PAHs are so low that their contribution, if there is some, to the $MFR_{\text{non-BC}}$ at 280 °C is minuscule. However, the results indicate that aerosol volatility is connected to anthropogenic influence over the whole measurement period. It was also observed that the BCF correlated positively with the PAH mass fraction with a correlation coefficient of 0.64 and a p-value of 0.03. This result suggests that motor vehicle emissions or other anthropogenic emissions, e.g. from residential heating, are the main source of BC in Hyytiälä area.

4.5 Indications of the chemical composition of $MFR_{\text{non-BC}}$

AMS data were available for two month-long campaigns: one in May–June and one in September–October 2008 (Fig. 1). The aerosol chemical composition as determined from the AMS is presented in Fig. 9 (upper panel). The $MFR_{\text{non-BC}}$ at 280 °C was compared to these data (Fig. 9, lower panel). It was assumed that the $MFR_{\text{non-BC}}$ at 280 °C is comparable with the AMS compounds evaporating at 600 °C. During the spring campaign the $MFR_{\text{non-BC}}$ at 280 °C did not have a significant correlation with the mass fractions of any of the AMS-compounds (Fig. 9, lower panel). AMS measurements during fall gave more insight to the chemical composition of the non-volatile (at 280 °C) residual. Although the median $MFR_{\text{non-BC}}$ at 280 °C was small (Fig. 5), the maximum value for the $MFR_{\text{non-BC}}$ at 280 °C was as high as 0.25. In the following we will concentrate on the comparisons between the VDMPS data and the AMS data from the fall campaign.

The VDMPS was temperature-scanning during the AMS campaigns. This made it possible to study the correlation between the aerosol chemical composition and the $MFR_{\text{non-BC}}$ at six different temperatures (see Fig. 10 for results from the fall campaign). Ammonia and sulfates correlated positively with the $MFR_{\text{non-BC}}$ at 280 °C only at temperatures below

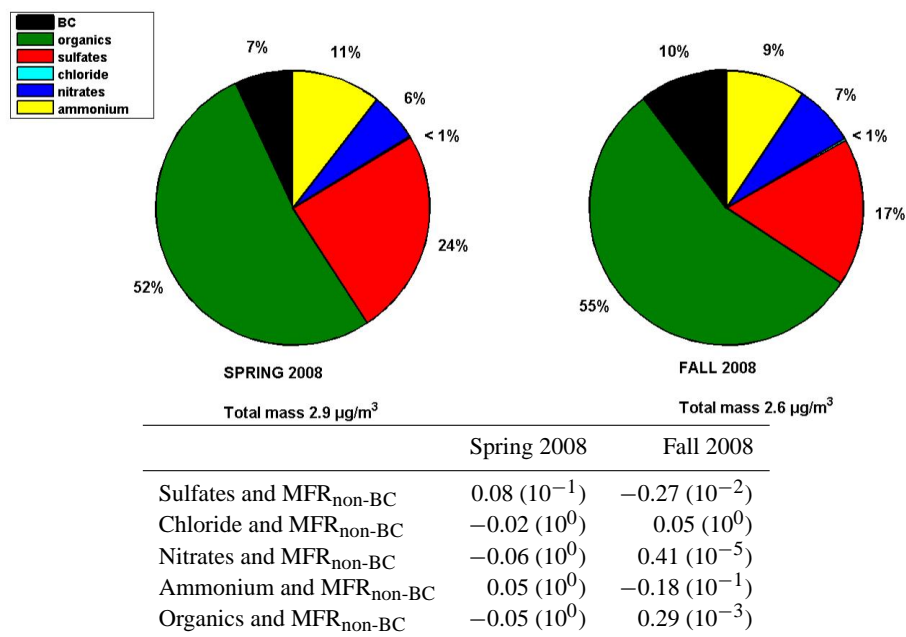


Fig. 9. Median aerosol chemical composition during AMS measurement campaigns in spring and fall 2008 is illustrated in the upper panel. BC mass measured with an aethalometer is included in the total aerosol mass and the median total mass concentrations of particles are provided. The correlation (r and p -value in parenthesis) between MFR_{non-BC} at 280 °C and mass fractions of AMS-compounds is presented in the lower panel.

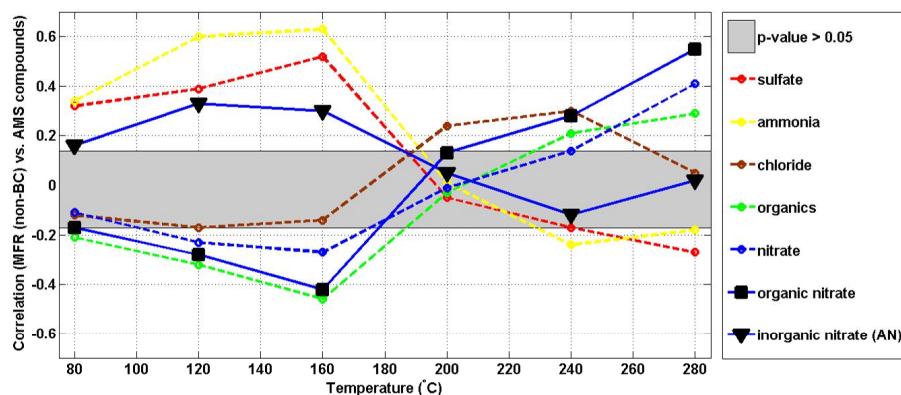


Fig. 10. Correlation coefficients between MFR_{non-BC} at 280 °C and mass fractions of AMS-compounds at six different temperatures. Mass concentrations (/fractions) of ammonium nitrate (AN) and organic nitrate were estimated from the AMS data. The grey area presents the correlation range ($r = -0.17 \dots 0.14$) where the correlations start to approach an area of not being statistically significant (p -value > 0.05).

160 °C. This is in line with previous studies reporting evaporation of ammonium sulfate at approximately 200 °C. Nitrates and organics, however, had a statistically significant positive linear correlation with the MFR_{non-BC} at 280 °C. Also, a positive correlation between MFR_{non-BC} and AMS chloride was observed at high heating temperatures. However, the mass loadings of chloride compounds observed by the AMS suggest the amount of chloride compounds is almost negligible in Hyytiälä (see Fig. 9 upper panel), and the amounts observed are most likely insufficient to have any effect on the MFR_{non-BC} within the measurement inaccuracies. During the

spring campaign the ammonium-sulfate behavior was similar to the fall campaign, whereas with organics and nitrates no clear correlation was found.

The relative importance of inorganic vs. organic nitrate was estimated by dividing the available ammonium to sulfate and the residual ammonium to nitrate. The remaining nitrate was assumed to be organic. All the nitrate was not bound to ammonium and thus, a notable amount of organic nitrate was also present (see Fig. A1). The correlations between the MFR_{non-BC} at 280 °C and the mass fractions of ammonium nitrate and organic nitrate are illustrated in Fig. 10. It can be

observed that the $\text{MFR}_{\text{non-BC}}$ at 280 °C correlated positively with the mass fraction of organic nitrate at high heating temperatures whereas the correlation behavior of ammonium nitrate was similar to ammonium and sulfate. While the correlations between the AMS mass fractions and the $\text{MFR}_{\text{non-BC}}$ at 280 °C were clear, it should be borne in mind that our analysis has been restricted to the total aerosol mass (20–500 nm), as no size-dependent AMS data was available. It is therefore not possible to say how the composition of the very low-volatile material is likely to depend on particle size or whether its composition is rather uniform throughout the size distribution.

In the fall 2008 particles arriving at Hyytiälä from the clean sector (i.e. from the forest, Tunved et al., 2006) had relatively more organic nitrates and very low-volatile material in them (Fig. A2, lower panel) as compared with other air mass directions. During spring such a correlation was not observed (Fig. A2, upper panel). Our results therefore suggest that the $\text{MFR}_{\text{non-BC}}$ at 280 °C is connected to organic nitrates in the fall period with clean air masses. This conclusion is, however, speculative and additional studies are needed to verify the presence of very low-volatile organonitrates.

Raatikainen et al. (2010) observed that aerosol nitrates, sulfates and ammonium correlated notably with low-volatile oxidized organic aerosol (LV-OOA) during spring 2005. LV-OOA was characterized as highly aged organic aerosol with an anthropogenic origin. It is possible that some of these LV-OOA compounds may not evaporate at 280 °C due to e.g. organic polymerization (Kalberer et al., 2004) or salt formation (Smith et al., 2010). However, the existence of very low-volatile LV-OOA was not clearly indicated by their results. It should be noted, however, that Raatikainen et al. (2010) investigated aerosol volatility in Hyytiälä over a shorter time period and with different instruments than in this study.

Our results, together with the findings of Raatikainen et al. (2010) clearly indicate that the processes controlling the composition of the remaining material in Hyytiälä vary depending on the season and probably no single explanation for the non-volatile residual at 280 °C can be found.

5 Summary and conclusions

Evaporation of 20–500 nm aerosol particles was investigated in a boreal forest site in Hyytiälä at the SMEAR II station during a two and a half year period from January 2008 to May 2010. The long-term measurements of aerosol volatility as well as additional data from the SMEAR II station made it possible to examine the seasonal behavior of the aerosol when heated, as well as to compare the data set with atmospheric parameters and with direct aerosol composition measurements. The higher the heating temperature was the more aerosol material evaporated, the temperature response of the evaporation being similar as in previous studies. At 80 °C

86 % of total aerosol mass remained after heating, whereas the mass fraction remaining (MFR) was 19 % at 280 °C.

It was found that the MFR varied with season. During winter months the MFR was the highest at all heating temperatures. In order to explain the existence of the MFR even at temperatures as high as 280 °C as well as its seasonal behavior, additional measurements, such as BC data, were included in the analysis. The maximum possible fraction of MFR at 280 °C explained by BC was dependent on the season: it was 55 % in the winter (median over December–February), 57 % in the spring (median over March–May), 87 % in the summer (median over June–August) and 75 % in the fall (median over September–November). Even though the aerosol BC fraction (BCF) was the highest during winter and fall months, during summer on some days all the non-volatile mass (at 280 °C) could have, in principle, been explained by BC. However, the correlation between the BCF and the MFR at 280 °C was the poorest during summer months – this could be partly due to small amount of data from summer months, but also probably linked to the contribution of other factors in defining the volatility of the aerosol in the summer as well.

Crustal material and sea salt are minor contributors to the total submicron aerosol mass in Hyytiälä, and the observed BC alone could not explain the existence of the material that did not volatilize even at 280 °C. Thus, roughly half of the aerosol non-volatile mass is explained by other compounds than BC, sea salt or crustal material. The properties of these non-BC compounds (forming the $\text{MFR}_{\text{non-BC}}$ at 280 °C) were investigated further by correlating the $\text{MFR}_{\text{non-BC}}$ at 280 °C with meteorological data and aerosol mass spectrometer data from spring and fall 2008. However, no single factor explaining the $\text{MFR}_{\text{non-BC}}$ at 280 °C was found.

Ambient temperature had a marked negative correlation with the $\text{MFR}_{\text{non-BC}}$ at 280 °C. It was also observed that high PAH aerosol mass fractions corresponded to high values of the $\text{MFR}_{\text{non-BC}}$ at 280 °C. These results suggest that the non-volatile aerosol fraction at 280 °C may contain also other compounds from anthropogenic sources in addition to BC. Positive correlations between the estimated organic nitrate and organic aerosol mass fractions from the AMS and the $\text{MFR}_{\text{non-BC}}$ at 280 °C were observed in fall 2008. These results suggest that very low-volatile organic compounds, e.g. polymerized organics or organonitrates, might exist in the least-volatile fraction of atmospheric aerosols not evaporating at 280 °C.

Our results indicate aerosol evaporation at 280 °C (with a residence time of 1.2 s) in a boreal forest environment cannot be explained with any single parameter – but rather is dependent on several environmental factors, varies with season and is affected by both anthropogenic and biogenic factors.

Appendix A

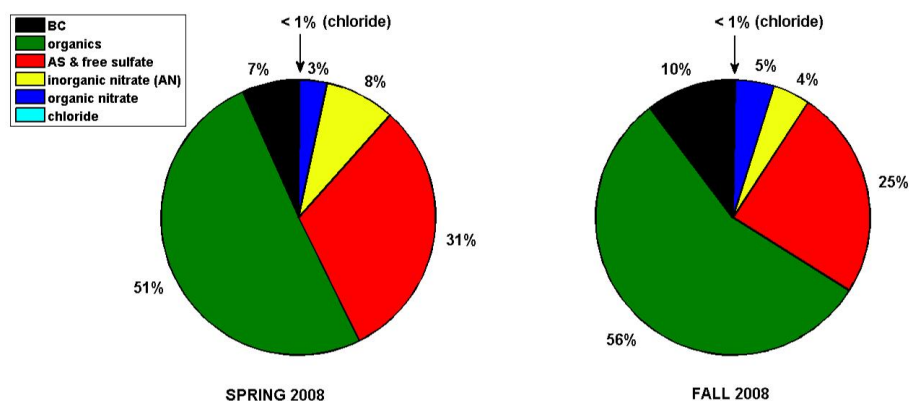


Fig. A1. Median aerosol chemical composition during AMS measurement campaigns in spring and fall 2008. Ammonium has been divided to sulfate and nitrate to form ammonium sulfate (AS) and ammonium nitrate (AN). Nitrate that is not neutralized by ammonium is assumed to be organic nitrate.

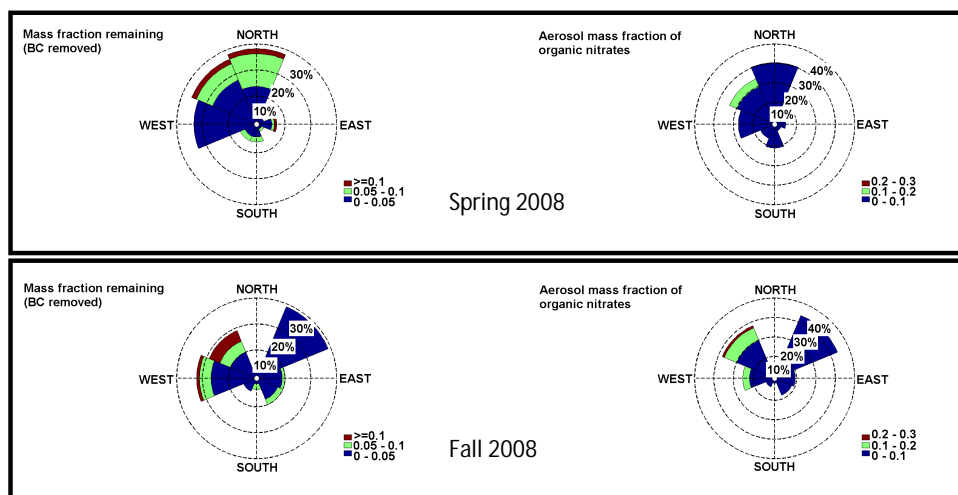


Fig. A2. Wind roses of aerosol mass fraction remaining ($MFR_{\text{non-BC}}$) after heating to 280 °C and the mass fraction of organic nitrates during AMS measurement campaign in spring 2008 (upper panel) and in fall 2008 (lower panel). The colors represent the mass fractions, and the percent values correspond to the amount of data from each wind direction bin. In fall 2008 aerosol organic nitrate mass fraction and $MFR_{\text{non-BC}}$ at 280 °C were correlating.

Table A1. Absorption mass cross-sections, $\sigma_{\text{abs},0}$ (from the manufacturer) and σ_{abs} (calculated based on Petzold and Schönlinner, 2004), and C parameters for correcting the BC concentrations measured with an aethalometer. C factors given are determined from Hohenpeissenberg data (Collaud-Coen et al., 2010) but because of the similar forest environment, these factors were assumed to describe the Hyytiälä aerosol alike.

Wavelength λ (nm)	$\sigma_{\text{abs},0}$ ($\text{m}^2 \text{g}^{-1}$)	σ_{abs} ($\text{m}^2 \text{g}^{-1}$)	C
370	39.53	11.36	2.64
470	31.12	8.95	2.75
521	28.07	8.07	2.81
590	24.79	7.13	2.87
660	22.16	6.37	2.93
880	16.62	4.78	3.09
950	15.39	4.43	3.13

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